Japanese Nationwide Study on the Association between Short-term Exposure to Particulate Matter and Mortality

Running title: Short-term exposure to PM$_{2.5}$ and mortality in Japan

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ABSTRACT

Background: From around 2012, the use of automated equipment for fine particulate matter (PM$_{2.5}$) measurement with equivalence to a reference method has become popular nationwide in Japan. This enabled us to perform a national health effect assessment employing PM$_{2.5}$ concentrations based on the standardised measurement method. We evaluated the association between non-accidental mortality and short-term exposure to PM$_{2.5}$ and coarse particulate matter (PM), with the latter estimated as the difference between suspended particulate matter and PM$_{2.5}$, for the fiscal years 2012-2014.

Methods: This was a time-stratified case-crossover study in 100 highly-populated Japanese cities. Mortality data was obtained from the Ministry of Health, Labour and Welfare. City-specific estimates of PM-mortality association were calculated by applying a conditional logistic regression analysis, and combined with a random-effects meta-analysis.

Results: The respective averages of daily mean concentration were 14.6 μg/m$^3$ for PM$_{2.5}$ and 6.4 μg/m$^3$ for coarse PM. A 10 μg/m$^3$ increase in PM$_{2.5}$ concentrations for the average of the day of death and the previous day was associated with an increase of 1.3% (95% confidence interval (CI), 0.9-1.6%) in total non-accidental mortality. For cause-specific mortality, PM$_{2.5}$ was positively associated with cardiovascular and respiratory mortality. After adjustment for PM$_{2.5}$, we observed a 1.4% (95% CI, 0.2-2.6%) increase in total mortality with a 10 μg/m$^3$ increase in coarse PM.

Conclusion: The study revealed that short-term exposure to PM$_{2.5}$ had adverse effects on total non-accidental, cardiovascular, and respiratory mortality in Japan. Coarse PM exposure also increased the risk of total mortality.

Key words fine particulate matter, coarse particle, mortality, case-crossover, Japanese
INTRODUCTION

The health effects of particulate matter, particularly fine particulate matter (PM$_{2.5}$ that passes through a size-selective inlet with a 50% cut-off level of 2.5 μm in aerodynamic diameter), are an international concern,$^1$ and epidemiologic evidence increases year by year. For instance, with respect to mortality as a classical outcome, a PM$_{2.5}$-daily mortality association has been established, and the effect size, on a national or multi-country scale, has been considered from the viewpoint of public health. One nationwide study, in 112 U.S. cities, revealed that an increase of 10 μg/m$^3$ in PM$_{2.5}$ exposure was associated with a 0.98% increase in all-cause mortality,$^2$ and another nationwide study, in 272 Chinese cities, estimated a similar increase of 0.22% in mortality.$^3$ Further, a multi-country study in 10 European cities estimated a 0.55% increase in mortality.$^4$ Nonetheless, evidence of the effects of PM$_{2.5}$ exposure on health in Japan is extremely limited, because the routine measurement of PM$_{2.5}$ in Japanese air pollution monitoring stations only began after the Air Quality Standard for PM$_{2.5}$ was established in 2009.$^5$ Previously, we reported that short-term exposure to PM$_{2.5}$ was positively associated with all-cause mortality in those aged 65 years or older in 20 regions between 2002 and 2004;$^6$ however, PM$_{2.5}$ concentration measurement equivalence was not ensured among the 20 automated measuring devices in the study.

To perform PM$_{2.5}$ concentration comparison among cities, systematic measurement error based on the measuring equipment should be avoided. Therefore, the Japanese Air Quality Standards have defined the PM$_{2.5}$ measurement method as follows: “mass measurement with filter sample collection which is designated as a reference method, or alternative automated methods, designated as equivalent methods, which are proved to have measurement performance comparable to the corresponding reference method”.$^5$ From around 2012, use of automated equipment enabling PM$_{2.5}$ measurement with equivalence to a reference method has become popular nationwide. This allowed us to perform a national health effect assessment which employed PM$_{2.5}$ concentrations based on the standardised
measurement method.

We thus conducted a multi-city case-crossover study of the association between short-term exposure to PM$_{2.5}$ and mortality in Japan. Additionally, with regard to exposure to coarse particulate matter (coarse PM), for which there is still limited health effects evidence globally, we explored the association with mortality using an exposure surrogate for coarse PM, estimated as the difference between the concentrations of suspended particulate matter (SPM that passes through a size-selective inlet with a 100% cut-off level of 10 μm in aerodynamic diameter) and PM$_{2.5}$.

METHODS

Study area

In the 2015 Japanese Population Census, there were 110 cities with a total population exceeding 200,000. Of these 110 cities, 99 had one or more ambient air pollution monitoring stations with automated PM$_{2.5}$ measuring equipment with measurement equivalence to the reference method. Additionally, we included, in the study area, Tsukuba city (population: roughly 227,000 people in the 2015 Population Census), where has no monitoring station but is the home of the Japan National Institute for Environmental Studies, by combining Tsukuba City and adjacent Tsuchiura City (which has a monitoring station) into one city. Our 100 target cities were located throughout the length of Japan from Hokkaido Prefecture (Sapporo, Asahikawa, and Hakodate Cities) to Okinawa Prefecture (Naha City). The population and size of the 100 target cities are presented in Supplementary table 1. The study protocol was reviewed and approved by the Ethics Committee of the National Institute for Environmental Studies (2017-004), and Faculty of Medicine, Toho University (A18051).

Environmental data
The hourly measurement of PM$_{2.5}$ concentrations by the reference method-equivalent automated PM$_{2.5}$ measuring equipment, measured at one ambient air pollution monitoring station within each city, from April 2012 to March 2015, were obtained from the Japan National Institute for Environmental Studies’ atmospheric environment database. The PM$_{2.5}$ measurement periods varied by city (Supplementary table 1). Among 100 cities, 66 cities had only one ambient air monitoring station measuring PM$_{2.5}$. According to the pollutant monitoring manual by the Ministry of the Environment, Japan, such station is located at the point that seems to be representative of pollutant concentrations in the community. For the remaining cities, which all had two or more ambient air pollution monitoring stations, we selected one that satisfied the following criteria: (1) the station measured other co-pollutants, including SPM, photochemical oxidants (Ox), nitrogen dioxide (NO$_2$), and sulphur dioxide (SO$_2$), in addition to PM$_{2.5}$; and/or (2) the station was located nearer to the centre of the city than other stations. However, in such cases, we confirmed a strong correlation between the PM$_{2.5}$ concentrations measured at the selected station and those of the other station(s) in the same city (mean correlation coefficient = 0.95). The locations of the monitoring stations in the 100 target Japanese cities are presented in Supplementary figure 1. The daily mean concentrations of PM$_{2.5}$ were calculated based on hourly measurements from 0 to 23 h, and no daily mean concentrations were calculated for days when more than four hourly measurements were missing. The median % missing days in the PM$_{2.5}$ data was 1.2% (interquartile range (IQR), 0.6-2.2%), and these days were excluded in the analysis.

The hourly measurements of other co-pollutants were also obtained, and the daily mean concentrations of SPM, NO$_2$ and SO$_2$, as well as the maximum 8-h concentrations of Ox, were calculated. In this study, the difference between the respective concentrations of SPM and PM$_{2.5}$ was used as an exposure surrogate for coarse PM. Internationally, coarse PM is defined as particulate matter measuring between 2.5 μm and 10 μm (PM$_{10-2.5}$) in diameter. However, the monitoring stations in Japan do not routinely measure PM$_{10}$ (particles that pass
through a size-selective inlet with a 50% cut-off level of 10 μm in aerodynamic diameter). For
the coarse PM-mortality association, we included 77 cities where concentrations of coarse PM
had been reasonably distributed.

Meteorological data (daily mean ambient temperature and relative humidity), measured
at the nearest meteorological observatory to each city, were collected from the Japan
Meteorological Agency. Finally, information on the weekly influenza counts was collected
from the Japan National Institute of Infectious Diseases, and an influenza epidemic was
considered to persist for the number of weeks above the 90 percentile of the distribution
during the study period.9

Mortality data
For all the study areas, we applied to the Ministry of Health, Labour and Welfare, Japan, for
use of the data from its Vital Statistic Survey, and obtained daily mortality records, including
the age and sex of the deceased, and the date and location of death, as well as the primary
cause of death according to the 10th edition of the International Classification of Diseases
(ICD-10). In this study, our primary outcome was total non-accidental death (ICD-10: A00
through R99), and specific causes were defined as: cardiovascular disease (I00 through I99);
coronary heart disease (I20 through I25); stroke (I60 through I69); and respiratory disease
(J00 through J99).

Statistical analysis
We used a time-stratified case-crossover design, and compared an individual’s exposure on
the “case” day with his or her exposure on the “control” days, when he or she did not become
a case.10 We defined the “case” day as the day of death, and selected three or four “control”
days from the same days of the week, in the same month and year, as the “case” day. For
instance, if an individual died on 15 Mar 2013, four control days were assigned: 1, 8, 22, and
29 March 2013. We excluded cases on national holidays (not include weekends), to avoid bias in selection control due to the difference between pollutant concentrations on holidays and on non-holidays.

To investigate the association between exposure to PM$_{2.5}$ and mortality, we first applied city-specific conditional logistic regression models. Based on the past studies, the choice of the lag0-1 (2-day moving average of the case-day and previous-day) concentrations of PM$_{2.5}$ was made before commencement of the study. The PM$_{2.5}$ concentrations at lag2 and lag3 were used to investigate the lag pattern of the PM$_{2.5}$-mortality association. We constructed a distributed lag model, simultaneously including PM$_{2.5}$ concentrations at lag0-1, lag2, and lag3, and adjusted for ambient temperature at lag0-1 using a five-knot natural cubic spline, relative humidity at lag0-1 using a three-knot cubic spline, and influenza epidemics. Odds ratios (ORs) with 95% confidence intervals (CIs) of mortality, based on an increase of 10 μg/m$^3$ in PM$_{2.5}$ concentration at lag0-1 were estimated, because a linear model was not inferior to a non-linear model with respect to estimation of the PM$_{2.5}$-mortality association.

Subsequently, the city-specific estimates were used to obtain pooled estimates on the PM$_{2.5}$-mortality association, through random-effects meta-analysis. The ORs were converted into excess risks approximated by [(OR - 1)×100]. The proportion of total variation in the city-specific estimates, attributable to heterogeneity, was assessed by computing the I$^2$ statistic.

The same two-stage approach was followed in estimating the association between coarse PM and mortality. We mutually adjusted for PM$_{2.5}$ and coarse PM after estimating each PM-mortality association in the single-pollutant model.

For sensitivity analysis, we constructed two-pollutant models; and another model additionally adjusted for temperature at lag2-21 to investigate the long delay of the cold effects. With respect to PM$_{2.5}$ exposure, to estimate the health effects at PM$_{2.5}$ concentrations not exceeding the daily World Health Organization (WHO) guideline values, we analysed after restricting to days with PM$_{2.5}$ concentrations at lag0-1 ≤ 25 μg/m$^3$. We also
performed stratified analyses by age (<75, ≥75 years) and sex, and examined seasonal variation in the PM-mortality association. Since the health effects of PM$_{2.5}$ seem likely to differ with the composition and sources of PM$_{2.5}$, we also explored region-specific associations: East Japan where domestic pollution is the major contributor to PM$_{2.5}$ concentrations; Central Japan; and West Japan where transboundary pollution makes a large contribution to PM$_{2.5}$ (Supplementary table 1). All analyses were performed using STATA version 14 (Stata Corporation, College Station, TX, USA).

RESULTS

In this nationwide analysis, the average daily mean PM$_{2.5}$ concentration was 14.6 (standard deviation = 8.3) μg/m$^3$ (Table 1), and the 98th percentile of daily mean concentrations was 37.8 μg/m$^3$. In terms of region-specific statistics, the respective averages of PM$_{2.5}$ concentration were 13.6 (8.0) μg/m$^3$ in East, 14.6 (8.3) μg/m$^3$ in Central, and 17.9 (9.5) μg/m$^3$ in West Japan. The average daily mean concentration of coarse PM was 6.4 (5.1) μg/m$^3$. The city-specific results for environmental factors are presented in Supplementary table 1. The Pearson’s correlation coefficients with PM$_{2.5}$ were 0.33 for coarse PM, 0.37 for Ox, 0.47 for NO$_2$, and 0.42 for SO$_2$ (Supplementary table 2). Among 1,347,152 non-accidental deaths, cardiovascular mortality accounted for 28.1% and respiratory mortality for 16.7% (Table 2).

When we explored the city-specific association between exposure to PM$_{2.5}$ and total non-accidental mortality, the point estimates of percentage increase for 10 μg/m$^3$ increase in PM$_{2.5}$ at lag0-1 showed a direction of increased mortality risk in 72 cities (Supplementary figure 2). We did not observe significant heterogeneity ($\chi^2$ statistic = 9.8%). Pooled estimates for the association between PM$_{2.5}$ exposure and mortality are presented in Table 3. Exposure to PM$_{2.5}$ at lag0-1 was associated with an increase in total mortality of 1.3% per 10 μg/m$^3$. 
The respective percentage increases for the association of exposure to
PM$_{2.5}$ and cause-specific mortality were 1.6 (95% CI, 0.8-2.4) for cardiovascular disease, 2.7
(95% CI, 1.0-4.4) for coronary heart disease, 1.3 (95% CI, −0.2 to 2.8) for stroke, and 1.5
(95% CI, 0.7-2.3) for respiratory disease. PM$_{2.5}$ at lag2 and lag3 was not associated with
either total or cause-specific mortality.

We performed sensitivity and stratified analyses of the association between PM$_{2.5}$
exposure and total non-accidental mortality (Figure 1). After adjustment for exposure to co-
pollutants, or for temperature at long lag, the percentage increase point estimates were
attenuated; however, the positive association between PM$_{2.5}$ and mortality persisted. Even
based on WHO guideline values, the total mortality increased by 1.1% (95% CI, 0.6-1.7) with
a 10 μg/m$^3$ increase in PM$_{2.5}$ at lag0-1. When we performed the stratified analyses, the point
estimates of percentage increase tended to be slightly greater in adults aged 75 years or older,
and in spring and autumn. The PM$_{2.5}$-mortality association varied somewhat by region: there
was a positive association in East Japan (percentage increase = 1.5, 95% CI, 0.9-2.0) and in
Central Japan (1.7, 1.1-2.3), but no clear positive association was found in West Japan (0.5,
−0.5 to 1.4).

We analysed the association between exposure to coarse PM and non-accidental
mortality in 77 cities (Supplementary figure 3), and summarise the results of pooling the city-
specific estimates in Table 4 and Supplementary table 3. The coarse PM concentrations at
lag0-1 were positively associated with an increased risk of total mortality (percentage increase
per 10 μg/m$^3$ = 2.3, 95% CI, 1.4-3.3) without heterogeneity (I$^2$ statistic = 5.5%). After
adjustment for exposure to PM$_{2.5}$, though the pooled estimate of percentage increase was
attenuated, the coarse PM-mortality association was positive (percentage increase = 1.4, 95%
CI, 0.2-2.6). In the two-pollutant model, which included both coarse PM and PM$_{2.5}$, the point
estimates of percentage increase for the association of coarse PM with cardiovascular and
respiratory mortality showed a positive direction, but were not statistically significant.
(percentage increase for cardiovascular mortality = 1.3, 95% CI, −0.9 to 3.5; and for
respiratory mortality = 0.9, −1.8 to 3.8). With regard to PM$_{2.5}$, adjustment for coarse PM did
not substantially affect the pattern of the PM$_{2.5}$-mortality association (Table 4).

For coarse PM-mortality association, the elevated risk was more evident in West Japan
than in East and Central Japan (Supplementary figure 4).

DISCUSSION

Under environmental conditions, in which the average (14.6 μg/m$^3$) and 98th percentile (37.8
μg/m$^3$) of daily mean PM$_{2.5}$ concentration were similar to those of the Japan Air Quality
Standard (annual mean ≤ 15 μg/m$^3$, daily mean ≤ 35 μg/m$^3$), an overall 10 μg/m$^3$ increase in
PM$_{2.5}$ concentrations on the day of death and the previous day was associated with a 1.3%
increase in total non-accidental mortality. This positive association was observed even when
we restricted to days with PM$_{2.5}$ concentrations at lag0-1 ≤ 25 μg/m$^3$. The adverse effects of
PM$_{2.5}$ in this study were consistent with the previous, 20-cities study in Japan, which observed
that a 10 μg/m$^3$ increase on the previous day was associated with a 0.88% increase in non-
accidental all-cause mortality for adults aged 65 years or older,$^6$ though direct comparison of
the effect size of the PM$_{2.5}$-mortality association was difficult due to the difference in the
number of cities, age category, and PM$_{2.5}$ measurement method involved.

Our effect estimate of PM$_{2.5}$ (a 1.3% increase in total mortality) tended to be higher
than the estimates in the studies of other countries: 0.98% (95% CI, 0.75-1.22%) in 112 U.S.
cities from 1999-2005,$^5$ 0.55% (0.27-0.84%) in 10 Mediterranean metropolitan areas from
2001-2010,$^4$ and 0.8% (0.3-1.2%) in a nationwide Dutch study from 2008-2009,$^{16}$ and 0.22%
(0.15-0.28%) in 272 Chinese cities from 2013-2015.$^3$ One reason for this may be partly
explained by differences in population structure. In Japan, the proportion of older population
tends to be higher than in other countries. As we observed, the effect estimate of PM$_{2.5}$ tended to be higher in those 75 years old or more, than in those less than 75 years old; and older individuals are likely more vulnerable to PM exposure. In addition, the difference may be influenced by regional differences in the composition and sources of PM$_{2.5}$. Even in Japan, though a positive direction in the association between PM$_{2.5}$ and mortality was observed in all three regions that were distinguished based on the composition and sources of PM$_{2.5}$, the mortality effect of PM$_{2.5}$ appeared to vary somewhat by region. In East Japan, domestic traffic-related pollution, as a source of elemental carbon (EC), made a major contribution to PM$_{2.5}$ concentrations; while sulfate (SO$_{4}^{2-}$) concentrations, influenced by transboundary pollution, were lower than in West Japan. Both EC and SO$_{4}^{2-}$ were associated with mortality, but then the largest association per unit mass was observed for EC. Such regional differences in the composition and sources of PM$_{2.5}$ may help interpret the apparent association in East Japan rather than in West Japan. Thus a related health effect assessment, focused on the composition and sources of PM$_{2.5}$, is required in future. Among the possible reasons for seasonal difference, spring and autumn tend to witness an increase in outdoor activities, and decrease in air conditioner usage, meaning increased exposure to outdoor air pollution.

We found that exposure to PM$_{2.5}$ was positively associated with both cardiovascular and respiratory mortality, and this was consistent with the results for the PM$_{2.5}$-mortality association in past studies. With respect to cardiovascular disease, we found that PM$_{2.5}$ exposure was associated with mortality due to coronary heart disease, but not clearly associated with stroke mortality. The existence of acute cardiac effects due to PM$_{2.5}$ exposure has been internationally accepted, and we believe that the supposition of a positive PM$_{2.5}$-stroke association is not far from gaining consensus. Indeed, there is steadily increasing evidence suggesting a positive association between short-term exposure to PM$_{2.5}$ and mortality due to stroke. In Tokyo, for example, PM$_{2.5}$ exposure was associated with an
elevated risk of elderly mortality related to stroke as well as coronary heart disease. In fact, the non-statistical elevated risk of stroke-related mortality we here observed may partially explain the fact that the acute fatality rate of coronary heart disease is higher than that of stroke.\footnote{28} 

The health effects of coarse PM are still under investigation.\footnote{8, 29} In the American multi-city study,\footnote{2} exposure to coarse PM, estimated as the difference between PM$_{10}$ and PM$_{2.5}$ (PM$_{10-2.5}$), was associated with an increase in all-cause mortality; however, no such association was observed in the European multi-city study.\footnote{4} In the present study, we used the difference between the respective concentrations of SPM and PM$_{2.5}$ as a surrogate marker of coarse PM, and observed a 1.4% increase in total mortality with a 10 μg/m$^3$ increase in coarse PM at lag0-1, after adjustment for PM$_{2.5}$. This finding was similar to the Japanese single-city (i.e., 23 wards of Tokyo) result for the association between coarse PM (defined as in our study) and all-cause mortality among adults aged 65 years or older between 2002 and 2013 (PM$_{2.5}$-adjusted percentage increase = 1.6%, 95% CI, 1.1-2.2%).\footnote{27} The clear association between coarse PM and mortality in West Japan might be partially related to the health effects of Asian dust originating in the deserts of China and Mongolia, which is mainly composed of coarse particles.\footnote{30-32} There were more days of Asian dust in West Japan than in East and Central Japan.\footnote{33} Our results did not show a clear association between coarse PM exposure and cause-specific mortality, but did suggest a positive direction in the association. Coarse particles cause inflammation with deposition in the bronchial tubes, but unlike fine particles, do not penetrate to the alveolus.\footnote{34, 35} Any significant conclusion regarding the health effects of coarse PM exposure in consideration of PM$_{2.5}$ exposure would require further evidence.

The main strength of the present study is that it provides the first nationwide evidence of an association between mortality and short-term exposure to PM$_{2.5}$ or coarse PM in Japan. Further, data for a large number of deaths was analysed, and the effects of PM exposure on mortality were estimated with narrow CIs. Another strength of the study is its use of PM$_{2.5}$
data measured by automated PM$_{2.5}$ measuring equipment with measurement equivalence to a
reference method. This enabled evaluation of the PM$_{2.5}$-mortality association without
allowance for systematic PM$_{2.5}$ measurement error due to the measuring equipment. On the
other hand, we should acknowledge some limitations in this study. There was random
measurement error in the exposure assessment, and coarse PM (difference between SPM and
PM$_{2.5}$) was more likely to be susceptible to such error than PM$_{2.5}$. Since this classical error
appears to be unrelated to outcome, we must interpret the results of the association between
coarse PM and mortality with allowance for an underestimate on the effect. As in past
environmental epidemiologic studies investigating acute health effects of air pollution,
another limitation of the study was the inevitability of exposure misclassification, giving that
the results were based on the use of pollutant data measured at a single monitoring station in
each city. It was difficult to estimate the biasing effect of a combination of this Berkson error
and classical error on our estimates of the PM-mortality association.$^{36}$

In conclusion, we observed that short-term exposure to PM$_{2.5}$ had adverse effects on
total non-accidental, cardiovascular, and respiratory mortality in Japan. Coarse PM exposure
also increased the risk of total mortality. Our findings suggest that we must continue
monitoring for exposure to PM, including the coarse component, and continue assessing the
health effects of such exposure.
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Competing Interests  None declared.
Reference


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Table 1. Overall statistics of environmental factors in 100 Japanese cities, April 2012 to March 2015

<table>
<thead>
<tr>
<th>Environmental factors (daily mean)</th>
<th>Number of cities</th>
<th>Mean (SD)</th>
<th>25th</th>
<th>50th</th>
<th>75th</th>
</tr>
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<tbody>
<tr>
<td>PM$_{2.5}$ ($\mu$g/m$^3$)</td>
<td>100</td>
<td>14.6 (8.3)</td>
<td>8.6</td>
<td>13.1</td>
<td>18.8</td>
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<tr>
<td>Coarse PM ($\mu$g/m$^3$)$^a$</td>
<td>77</td>
<td>6.4 (5.1)</td>
<td>2.9</td>
<td>5.6</td>
<td>8.9</td>
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<tr>
<td>Ox (ppb)$^b$</td>
<td>95</td>
<td>42.3 (15.2)</td>
<td>31.8</td>
<td>40.6</td>
<td>51.7</td>
</tr>
<tr>
<td>NO$_2$ (ppb)</td>
<td>99</td>
<td>12.4 (6.1)</td>
<td>7.9</td>
<td>11.2</td>
<td>15.7</td>
</tr>
<tr>
<td>SO$_2$ (ppb)</td>
<td>91</td>
<td>2.1 (1.2)</td>
<td>1.3</td>
<td>1.9</td>
<td>2.7</td>
</tr>
<tr>
<td>Ambient temperature ($^\circ$C)</td>
<td>100</td>
<td>15.7 (8.5)</td>
<td>7.7</td>
<td>16.0</td>
<td>23.0</td>
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<tr>
<td>Relative humidity (%)</td>
<td>100</td>
<td>67 (13)</td>
<td>58</td>
<td>68</td>
<td>76</td>
</tr>
</tbody>
</table>

NO$_2$, nitrogen dioxide; Ox, photochemical oxidants; PM, particulate matter; SD, standard deviation; SO$_2$, sulphur dioxide.

$^a$Concentrations of coarse PM were calculated by subtracting PM$_{2.5}$ concentrations from those of suspended particulate matter.

$^b$Daily maximum 8-h mean concentrations.
### Table 2. Overall characteristics of non-accidental deaths (1,347,152 deaths) in 100 Japanese cities, April 2012 to March 2015

<table>
<thead>
<tr>
<th>Outcome</th>
<th>Percent</th>
</tr>
</thead>
<tbody>
<tr>
<td>Total non-accidental (ICD-10: A00-R99)</td>
<td>100</td>
</tr>
<tr>
<td>Age (years)</td>
<td></td>
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<tr>
<td>&lt; 75</td>
<td>27.9</td>
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<tr>
<td>≥ 75</td>
<td>72.1</td>
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<td>Sex</td>
<td></td>
</tr>
<tr>
<td>Men</td>
<td>51.9</td>
</tr>
<tr>
<td>Women</td>
<td>48.1</td>
</tr>
<tr>
<td>Cardiovascular disease (ICD-10: I00-99)</td>
<td>28.1</td>
</tr>
<tr>
<td>Coronary heart disease (ICD-10: I20-25)</td>
<td>6.4</td>
</tr>
<tr>
<td>Stroke (ICD-10: I60-69)</td>
<td>9.2</td>
</tr>
<tr>
<td>Respiratory disease (ICD-10: J00-99)</td>
<td>16.7</td>
</tr>
</tbody>
</table>

ICD, international classification of diseases.
Table 3.  Pooled estimates for the association between PM$_{2.5}$ exposure and mortality in 100 Japanese cities, April 2012 to March 2015

<table>
<thead>
<tr>
<th>Outcome</th>
<th>Lag0-1 Percentage increase for 10 μg/m$^3$ increase in PM$_{2.5}$ (95% CI)</th>
<th>Lag2 Percentage increase for 10 μg/m$^3$ increase in PM$_{2.5}$ (95% CI)</th>
<th>Lag3 Percentage increase for 10 μg/m$^3$ increase in PM$_{2.5}$ (95% CI)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Total non-accidental (ICD-10: A00-R99)</td>
<td>1.3 (0.9 to 1.6)</td>
<td>0.2 (-0.2 to 0.5)</td>
<td>-0.3 (-0.7 to 0.1)</td>
</tr>
<tr>
<td>Cardiovascular disease (ICD-10: I00-99)</td>
<td>1.6 (0.8 to 2.4)</td>
<td>0.5 (-0.1 to 1.1)</td>
<td>-0.4 (-1.1 to 0.4)</td>
</tr>
<tr>
<td>Coronary heart disease (ICD-10: I20-25)</td>
<td>2.7 (1.0 to 4.4)</td>
<td>1.0 (-0.6 to 2.6)</td>
<td>1.0 (-0.1 to 2.2)</td>
</tr>
<tr>
<td>Stroke (ICD-10: I60-69)</td>
<td>1.3 (-0.2 to 2.8)</td>
<td>-0.3 (-1.4 to 0.8)</td>
<td>-0.3 (-1.5 to 0.9)</td>
</tr>
<tr>
<td>Respiratory disease (ICD-10: J00-99)</td>
<td>1.5 (0.7 to 2.3)</td>
<td>-0.3 (-1.3 to 0.8)</td>
<td>0.1 (-1.0 to 1.1)</td>
</tr>
</tbody>
</table>

CI, confidence interval; ICD, international classification of diseases; PM, particulate matter.

*Adjusted for PM$_{2.5}$ at lag0-1, lag2 and lag3 simultaneously, ambient temperature at lag0-1, relative humidity at lag0-1, and influenza epidemics.
Table 4. Pooled estimates for non-accidental mortality associated with PM$_{2.5}$ and coarse PM in 77 Japanese cities, April 2012 to March 2015

<table>
<thead>
<tr>
<th>Outcome</th>
<th>PM$_{2.5}$ Percentage increase for 10 μg/m$^3$ increase at lag0-1$^b$ (95% CI)</th>
<th>Coarse PM$^a$ Percentage increase for 10 μg/m$^3$ increase at lag0-1$^b$ (95% CI)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Single pollutant model</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Total non-accidental (ICD-10: A00-R99)</td>
<td>1.2 (0.8 to 1.7)</td>
<td>2.3 (1.4 to 3.3)</td>
</tr>
<tr>
<td>Cardiovascular disease (ICD-10: I00-99)</td>
<td>1.5 (0.6 to 2.4)</td>
<td>2.6 (0.7 to 4.4)</td>
</tr>
<tr>
<td>Coronary heart disease (ICD-10: I20-25)</td>
<td>1.8 (-0.2 to 3.9)</td>
<td>3.8 (0 to 7.7)</td>
</tr>
<tr>
<td>Stroke (ICD-10: I60-69)</td>
<td>1.0 (-0.6 to 2.7)</td>
<td>2.7 (-0.9 to 6.4)</td>
</tr>
<tr>
<td>Respiratory disease (ICD-10: J00-99)</td>
<td>1.5 (0.6 to 2.4)</td>
<td>1.4 (-0.8 to 3.7)</td>
</tr>
<tr>
<td>Two pollutant model$^c$</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Total non-accidental (ICD-10: A00-R99)</td>
<td>1.2 (0.6 to 1.7)</td>
<td>1.4 (0.2 to 2.6)</td>
</tr>
<tr>
<td>Cardiovascular disease (ICD-10: I00-99)</td>
<td>1.4 (0.3 to 2.4)</td>
<td>1.3 (-0.9 to 3.5)</td>
</tr>
<tr>
<td>Coronary heart disease (ICD-10: I20-25)</td>
<td>2.5 (0.3 to 4.6)</td>
<td>2.5 (-1.7 to 6.7)</td>
</tr>
<tr>
<td>Stroke (ICD-10: I60-69)</td>
<td>0.6 (-1.0 to 2.3)</td>
<td>2.9 (-0.9 to 6.9)</td>
</tr>
<tr>
<td>Respiratory disease (ICD-10: J00-99)</td>
<td>2.1 (1.0 to 3.2)</td>
<td>0.9 (-1.8 to 3.8)</td>
</tr>
</tbody>
</table>

CI, confidence interval; ICD, international classification of diseases; PM, particulate matter

$^a$Concentrations of coarse PM were calculated by subtracting PM$_{2.5}$ concentrations from those of suspended particulate matter.

$^b$Adjusted for PM$_{2.5}$ or coarse PM at lag2 and lag3, ambient temperature at lag0-1, relative humidity at lag0-1, and influenza epidemics.

$^c$We included both PM$_{2.5}$ and coarse PM in the model.
Figure legend

Figure 1. Sensitivity and stratified analyses of the association between PM$_{2.5}$ exposure and total non-accidental mortality in 100 Japanese cities, April 2012 to March 2015. CI, confidence interval; ICD, international classification of diseases; NO$_2$, nitrogen dioxide; Ox, photochemical oxidants; PM, particulate matter; SO$_2$, sulphur dioxide

*Adjusted for PM$_{2.5}$ at lag2 and lag3, ambient temperature at lag0-1, relative humidity at lag0-1, and influenza epidemics.

*bThe sum of the region was 99, because we did not include Naha, Okinawa, in West Japan due to geographical differences (see Supplementary table 1).
**Supplementary information**

**Supplementary table 1.** Descriptive data on the study period, total non-accidental death, and environmental factors for 100 Japanese cities

**Supplementary table 2.** Pearson's correlation coefficients among daily mean pollutant concentrations

**Supplementary table 3.** Pooled estimates for the association between coarse PM exposure and mortality in 77 Japanese cities, April 2012 to March 2015

**Supplementary figure 1.** Location of monitoring stations in 100 Japanese cities

**Supplementary figure 2.** City-specific odds ratios (ORs) of total non-accidental death with a 10 μg/m³ increase in PM2.5 concentrations at lag0-1

**Supplementary figure 3.** City-specific odds ratios (ORs) of total non-accidental death with a 10 μg/m³ increase in coarse particulate matter concentrations at lag0-1

**Supplementary figure 4.** Sensitivity and stratified analyses of the association between coarse PM exposure and total non-accidental mortality
Figure 1

Percentage increase for 10 µg/m³ increase in PM$_{2.5}$ at lag0-1$^a$

<table>
<thead>
<tr>
<th>Category</th>
<th>Number of cities</th>
</tr>
</thead>
<tbody>
<tr>
<td>Total non-accidental (ICD-10: A00-R99)</td>
<td>100</td>
</tr>
<tr>
<td>Two-pollutant model</td>
<td></td>
</tr>
<tr>
<td>Adjusted for Ox at lag0-1</td>
<td>95</td>
</tr>
<tr>
<td>Adjusted for NO$_2$ at lag0-1</td>
<td>99</td>
</tr>
<tr>
<td>Adjusted for SO$_2$ at lag0-1</td>
<td>91</td>
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<tr>
<td>Adjusted for temperature at lag2-21</td>
<td>100</td>
</tr>
<tr>
<td>Restricting to lag0-1 ≤ 25 µg/m³</td>
<td>100</td>
</tr>
<tr>
<td>Age (years)</td>
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<tr>
<td>&lt; 75</td>
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<tr>
<td>≥ 75</td>
<td>100</td>
</tr>
<tr>
<td>Sex</td>
<td></td>
</tr>
<tr>
<td>Men</td>
<td>100</td>
</tr>
<tr>
<td>Women</td>
<td>100</td>
</tr>
<tr>
<td>Season</td>
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</tr>
<tr>
<td>Spring (Mar to May)</td>
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<tr>
<td>Summer (Jun to Aug)</td>
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<tr>
<td>Autumn (Sep to Nov)</td>
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</tr>
<tr>
<td>Winter (Dec to Feb)</td>
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<tr>
<td>Region$^b$</td>
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<tr>
<td>East</td>
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<tr>
<td>Central</td>
<td>32</td>
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<tr>
<td>West</td>
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